

Understanding the spatio-temporal variability of fisheries data for better bottom trawling management practices in the Catalan margin (NW Mediterranean Sea)

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ABSTRACT

The long-time fishery tradition on the Mediterranean coastal region without a biological-base management strategy has led to an overexploitation of the main fishing resources. To overturn this trend, the European Union implemented the Western Mediterranean Multi-Annual Plan (WMMAP) aiming to better manage key commercial species caught by bottom trawling. The data to drive management decisions should be obtained from fisheries monitoring programs designed to collect data that reflect dynamics of both biological and spatio-temporal trends in different spatial extents and scales. With this in mind, the aim of this research is to analyze the spatio-temporal variability of the commercial fraction of the catch, focusing on the main fishing resources of the Catalan bottom trawling fleet throughout the GSA 6. The Institut Català de Recerca per a la Governança del Mar (ICATMAR) designed a locally-based monitoring program to collect biological data for seven main key fishing resources of the area including the spottail mantis shrimp, the red mullet, the horned octopus, the European hake, the deep-water rose shrimp, the Norway lobster and the blue and red shrimp. The study shows the patterns that affect the main fishing resources exploited by the Catalan bottom trawling fleet over space and time through the study of biological data relevant to fisheries. To develop best management practices, we suggest that this type of sampling be included in Mediterranean areas to complement the current Data Collection Framework (DCF) program, so that fisheries can be managed evaluating the biology of the species along with the social component of the fisheries structure.

1. Introduction

In the Mediterranean Sea, maritime activities and the exploitation of marine living resources have remained relevant for thousands of years [1], and this long-time tradition has led to overfishing [2]. In response to this, in 2000 the European Union Data Collection Framework (DCF) established for member states the obligation to collect, manage and annually report biological, environmental and socioeconomic fisheries

data as a source of scientific advice for management purposes (EU 2017/1004).

Fisheries data collection in the Mediterranean Sea is currently carried out through both fisheries-dependent and -independent information, collected by Geographical Subareas (GSAs), the management units defined by the General Fisheries Commission for the Mediterranean of the United Nations Food and Agriculture Organization (Resolution GFCM/33/2009/2). The geographical subareas are often used to define

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the size of the stock unit, meaning that the advice derived from stock assessment is referred to a particular GSA or a combination of GSAs. In practice, neither the distribution nor the status of a given stock are considered at a scale smaller than the GSA. In the GSA 6, the NE Spanish Mediterranean coast, bottom trawling fisheries-dependent data are gathered through the monitoring of commercial fisheries data on board and in auctions of some ports along the GSA 6, where the area is considered as a single unit. Bottom trawling fisheries-independent data are collected in annual surveys (MEDITS), in place since 1994 in late spring months, with an average of 85 yearly hauls along the whole GSA 6 [3,4]. Although these programs obtain accurate and reliable data on population structure and fishing stocks, the extension of the management unit is wide, encompassing an area with the highest number of fishers' associations and fishing vessels among all Mediterranean countries, so that aiming for an exhaustive sampling or year-round data collection along this whole area would not be feasible.

The design of a common management strategy for the EU is a complex matter, since fishing activity patterns and socio-economic contexts are diverse [5,6]. The management strategies at present in place are typically designed based on the North Atlantic fishery model, i.e., large vessels belonging to a few big companies, wide fishing areas trawled for weeks at a time, few fishery associations per region with non-daily auctions and large management units that suit a homogenous socio-economic structure. As a consequence, these strategies do not always translate well to the Spanish Mediterranean context, based on medium-small, family-owned vessels, with an average power of 180 kW, gross tonnage (GT) of 58, and length overall (LOA) of 20 m (versus 1257 kW, 1182 GT, and 60 m LOA for Spanish bottom trawl vessels fishing in the North Atlantic, respectively; EC Fleet Register 2024).

Another main characteristic of Mediterranean fisheries is the use of multiple gears and their markedly multi-specific nature, including a broad variety of target and discard species [7,8]. In fact, fisheries management in the Western Mediterranean Sea has traditionally followed a single-species approach (EU Reg. 2019/1022) [9,10], but some authors have argued that applying these single-management strategies to multi-specific fisheries may not be a good fit [11,12]. Finally, Mediterranean fishing activities also exhibit great variations from one area to another, from the point of view of production methods as well as the adaptation of human communities to the physical and biological environmental conditions [2].

Within the NW Mediterranean Sea, the Catalan coast comprises the northern half of the GSA 6, between the Spanish-French border and the Ebre Delta. Fisheries activities are deeply rooted in Catalan culture [13] and have historically been a main source of income and cultural identity for coastal communities. The three different modalities of commercial fishing practiced in the area are bottom trawling, purse seine fishing, and a large variety of small-scale fisheries. In 2022, the highest yearly revenues, around 55 million € (60 % of the total [14]), were reported by the bottom trawling fleet. The importance of bottom trawling spurred the development of a strong legislation that has already been in place for decades. The main management measures set by the 1960s pioneer collaborative management plan "Pla Castelló" – i.e. the widening of the mesh size up to 40 mm, and restrictions on number, fishing capacity and navigation time of vessels – are still enforced by the Spanish fishery regulation with a few updates after present EU regulations (EC 1626/1994; EC 1967/2006; Real Decreto 1440/1999). Along the Catalan coast, the collaboration among stakeholders has continued since, resulting in other self-regulated management initiatives [e.g. 14]. In this sense, each port may have its own implementation of fisheries regulations by internal agreements that go beyond European or member states management legislation. Actually, some of them host local co-management plans [15] focused on limiting fishing effort and enforcing biological monitoring of the target species and the communities (e.g., the blue and red shrimp in Palamós) [16]. These particularities call for a richer set of data that can inform decisions at a finer scale. The consideration of diverse types of data such as satellite-based

information of fishing activities and of primary production, combined at a fine spatial scale, could improve the understanding of the ecosystems' response to environmental driving forces, and of the main trends of fishery indicators such as yield, productivity and overexploitation rate of fishing stocks [17].

In 2018 the EU regulation COM/2018/0115 final - 2018/050 (COD) implemented the Western Mediterranean Multi-Annual Plan (WMMAP), which establishes a progressive reduction in fishing effort (in days) for the bottom trawling fleet, in an effort to attain sustainable fishing securing economic, employment and social benefits based on the management of the following key commercial species: the red mullet, the European hake, the deep-water rose shrimp, the Norway lobster, and the blue and red shrimp. In Catalonia, and following the EU DCF, the Directorate-General for Maritime Policy and Sustainable Fisheries of the Catalan Government and the Institut de Ciències del Mar (ICM-CSIC) promoted the *Institut Català de Recerca per a la Governança del Mar* (ICATMAR), an autonomous organization whose main goal is to generate scientific advice to regional, state and European administrations for management purposes. Since 2019, ICATMAR has developed and implemented a fisheries monitoring program that aims to complement the data collection protocol already in place, working at a local scale to gather information on spatial and seasonal variability of the factors that can shape fishing stocks dynamics [4]. The program monitors the main target species of the Catalan commercial fleet but the present study focuses only on the bottom trawling fleet because of its key importance in landings and revenues of the northern GSA 6.

This study aims to analyze the spatio-temporal variability of the main fishing resources of the Catalan bottom trawling fleet throughout the GSA 6. For this purpose, we analyzed biological data collected by a locally-based monitoring for seven key species of the area: the spottail mantis shrimp, the red mullet, the horned octopus, the European hake, the deep-water rose shrimp, the Norway lobster, and the blue and red shrimp. These data could potentially complement the official EU fisheries monitoring system dataset (DCF) to enrich the stock assessment models and improve accuracy in management decision-making processes.

2. Materials and methods

2.1. Study area

The study area comprises the entire Catalan coast, with 580 km of coastline, constitutes the northern half of the GSA 6, which in whole comprises from Cape Creus down to Cartagena (Resolution GFCM/33/2009/2). (Fig. 1A). The spatial distribution of the fishing effort by port shows the local-scale structure of the Catalan bottom trawling sector: the fleet of each port fishes in waters directly off their port, and overlapping among influence areas of the different ports is low (around 10 % in fishing hours per km², Mingote et al., in prep) and exists only among adjacent ports (Fig. 1A). We chose the 9 ports with the highest annual revenues which are also equidistant throughout the territory (Fig. 1A), attending to both hydrographic and geomorphological characteristics of the continental margin (shelf and slope) [18]. In the northernmost part of the Catalan coast, the continental shelf is cut by deep submarine canyons that allow for deep-sea fisheries activity while remaining relatively close to the coast. The coast between Arenys de Mar and Tarragona is characterized by the absence of submarine canyons and the influence of densely populated areas such as Barcelona, Vilanova i la Geltrú or Tarragona. The southernmost area of the coast is deeply affected by the discharge of the river Ebre, which generates a delta between L'Ampolla and La Ràpita. The division of the Catalan coast in these three zones responds to the morphometric mesoscale analysis of the margin [19] and has been used in previous faunistic assemblage studies [20]. Accordingly, the ports were classified in three sampling zones (north, center and south) as follows: Roses, Palamós, Blanes and Arenys de Mar in the northern zone; Barcelona, Vilanova i la Geltrú and

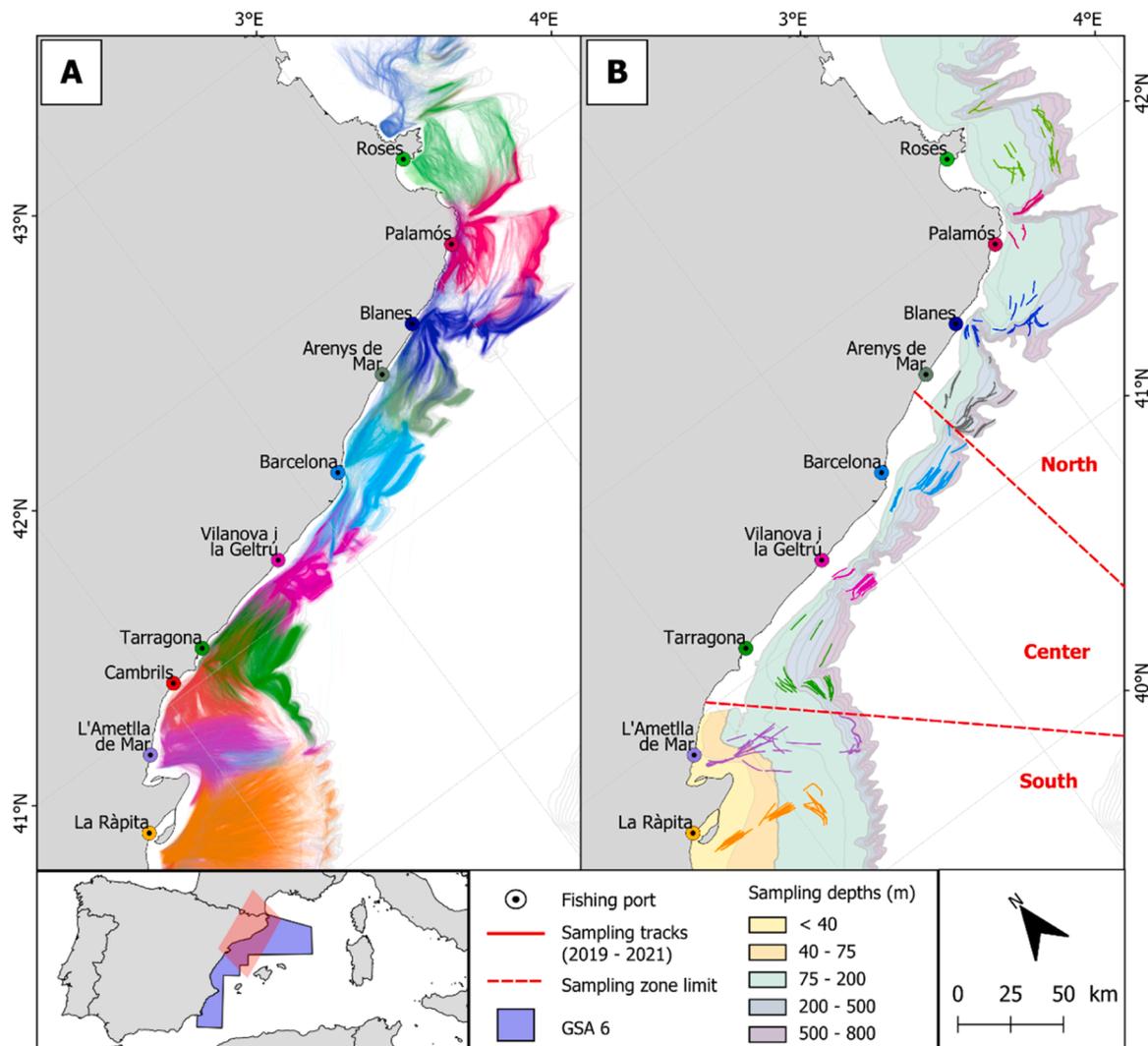


Fig. 1. Maps of the study area. The bottom image shows the GSA 6 (blue area), and the Catalan coast (red area). A: In color, fishing tracks (from Vessel Monitoring System, VMS data; for further information see [21]) of the Catalan bottom trawling fleet over the year 2021 represented by port. B: Sampling tracks followed in the ICATMAR monitoring in 2021, represented by zone (north, center, south), port and sampling depth range.

Tarragona in the central zone; and L'Ametlla de Mar and La Ràpita in the southern zone (Fig. 1B).

As described in the EU Reg. 1967/2006, the trawling fishing fleet from the Catalan coast is allowed to fish between 50 m depth (or 3 nautical miles from the coast) and 1000 m depth, five days per week (on weekdays) with a maximum of 12 hours per day (EC 1626/1994; EC 1967/2006). In shallow areas, such as the Ebre Delta (southernmost part of this study area; South Catalonia), trawling is allowed beyond 3 nautical miles off the coast, regardless of bottom depth.

The spatial distribution of the fishing effort shows high fishing activity in all depths where the bottom trawl fishery is allowed to fish by regulation. However, specific fishing effort varies according to the spatial distribution of the target species. Accordingly, different sampling depth ranges were defined along the Catalan coast. In the north and center zones, the sampling depth ranges defined are deep shelf (75 – 200 m), upper slope (200 – 500 m), and lower slope (500 – 800 m). In the southern area, which has a distinct geomorphological structure and bottom sedimentary characteristics, the defined sampling depths are coastal delta shelf (< 40 m), middle delta shelf (40 – 75 m), and deep shelf (75 – 200 m; Fig. 1). These depth strata are very similar to those applied by the MEDITS protocol, making the datasets potentially comparable [22].

2.2. Field sampling

Sampling trips were carried out on board trawling fishing vessels, with no change to their usual fishing activity, over commercial fishing grounds. Each of the three sampling zones was sampled monthly, with a rotation among the three ports of each zone, resulting in one quarterly sampling at each of the 9 selected ports. This allowed for the monitoring of seasonal patterns throughout the year. Each sampling day included three experimental hauls on board the same vessel, each one at one of the three depth ranges previously explained, within the high-fishing effort areas of each port (Supplementary Figure 2). Exceptionally, in 2019 and 2020, seven of the Southern zone hauls were conducted in the lower slope. The average depth for each haul was then estimated by calculating an average point between the start and end points of each haul. Mesh size was 40-mm square for all hauls except in Palamós lower slope, where the self-enforced local management measures for the blue and red shrimp fishery require a 50-mm squared mesh (BOE 2018, APM/532/2018). Each haul was GPS-recorded with a start and end point, fishing time and gear width. These measurements were used to calculate the total swept area per haul in order to standardize species biomass and abundance values [21]. The sampling is designed to run a complete cycle throughout a calendar year so that the data are comparable annually, and the data analyzed in this study correspond to the years

2019–2021.

2.3. Analysis of the catch

The fishers sort the catch into two categories: commercial, i.e., individuals of commercial species to be sold in the fish auction, and discarded, i.e., all other organic and inorganic organisms and items. The discarded fraction includes individuals of non-commercial species and/or undersized or damaged individuals of commercial species – as well as natural debris and marine litter. All species of the commercial fraction, including fish, crustaceans and cephalopods among others, are identified and measured on board (total length, cephalothorax length and mantle length, respectively). From all the species identified, seven target species were chosen. Five are species of interest featured in the WMMAP regulation (EU Reg. 2019/1022) – the European hake (*Merluccius merluccius*), the red mullet (*Mullus barbatus*), the blue and red shrimp (*Aristeus antennatus*), the Norway lobster (*Nephrops norvegicus*), the deep-water rose shrimp (*Parapenaeus longirostris*) – whereas the spottail mantis shrimp (*Squilla mantis*), and the horned octopus (*Eledone cirrhosa*), while not mentioned in the regulation, are relevant in the economy of the area [17].

A subsample of at least 30 individuals of each target species is preserved in coolers and transported to the laboratory. Then, the individuals are stored at 4°C and processed the following day of the sampling. All individuals are measured and weighed, and their reproductive stage was assessed. Different biological data are acquired depending on the target species, detailed in [Supplementary Table 1](#). For further information on the measurements taken on the other studied samples not reported in this study, see Annex I. The ICATMAR website (icatmar.cat) offers further information on all the studies done since 2019 and a data viewer with public and easy-access fisheries data [21].

2.4. Statistical analyses

For the purposes of this study, three different datasets were analyzed separately: all commercial species composition data, For the purposes of this study, three different datasets were analyzed separately: all commercial species composition data, and length frequency distribution, abundance and biomass of the target species.

The biomass of the commercial fraction of our sampling was calculated and standardized per square kilometer trawled. Data were square-root transformed in order to reduce the effect of extreme values, and the Bray-Curtis similarity index was calculated among hauls. Then, the data were represented in Non-Metric Multidimensional Scaling using the package *vegan* in R software [23], and statistically tested using PERMANOVA in R package *vegan* with 999 permutations. The three factors considered were season, zone, and depth. Pairwise comparisons between seasons, zones, and depths were also tested by subsetting the pair of factors to compare, where the function *p.adjust* (Bonferroni method) was applied in order to correct p-value of multiple comparisons. A similarity percentage analysis (SIMPER) was used to detect the species accounting for significant differences between depths.

Differences in the length frequency data obtained in our samplings for the seven target species were analyzed using a Kolmogorov-Smirnoff test by year, season, zone, and depth, with the R package *fishmethods* [23]. Only data from depths where each species is distributed were selected for the analyses (see [Table 1](#)) [24].

Differences in the abundance (individuals·km⁻²) and biomass (kg·km⁻²) for the seven target species were analyzed among years, seasons, areas, and depths using Generalized Linear Models (GLM), with the general formula as follows:

$$\text{abundance} \sim \text{year} + \text{season} + \text{zone} + \text{depth}$$

$$\text{biomass} \sim \text{year} + \text{season} + \text{zone} + \text{depth}$$

Data were not transformed and outliers were removed in a prior data validation process. Due to the zero-inflated nature of our data, the selected family distribution was negative binomial. The quantile-

Table 1

Number of individuals sampled every year for the seven target species, number of hauls where individuals of each species were analyzed, and depth strata where the species were distributed.

| Species | Year | Number of individuals sampled | Number of hauls | Depth strata |
|---------------------------------|------|-------------------------------|-----------------|--------------------|
| <i>Squilla mantis</i> | 2019 | 1862 | 18 | Coastal |
| | 2020 | 942 | 12 | delta shelf |
| | 2021 | 1500 | 14 | Middle delta shelf |
| <i>Mullus barbatus</i> | 2019 | 2181 | 56 | Coastal |
| | 2020 | 1397 | 41 | delta shelf |
| | | | | Middle delta shelf |
| <i>Eledone cirrhosa</i> | 2019 | 910 | 68 | Deep shelf |
| | 2020 | 696 | 58 | Coastal |
| | 2021 | 644 | 89 | delta shelf |
| <i>Merluccius merluccius</i> | 2019 | 3849 | 116 | Middle delta shelf |
| | 2020 | 2736 | 78 | Deep shelf |
| | 2021 | 4563 | 90 | Upper slope |
| <i>Parapenaeus longirostris</i> | 2019 | 3324 | 63 | Lower slope |
| | 2020 | 3527 | 53 | Coastal |
| | 2021 | 4491 | 62 | delta shelf |
| <i>Nephrops norvegicus</i> | 2019 | 6166 | 46 | Middle delta shelf |
| | 2020 | 4343 | 45 | Deep shelf |
| | 2021 | 5011 | 42 | Upper slope |
| <i>Aristeus antennatus</i> | 2019 | 9185 | 34 | Lower slope |
| | 2020 | 7650 | 28 | |
| | 2021 | 7336 | 28 | |

quantile (Q-Q) and Residuals vs. Predicted plots were explored through the package *DHARMA* to check for compliance of statistical assumptions and model fit. An analysis of the variance (ANOVA) was performed on the results of the model to clarify the significance of each factor. Then, pairwise comparisons were conducted for each pair of levels of each factor. The GLM analysis was done with the package *MASS* and the pairwise comparisons with *emmeans* [23].

3. Results

3.1. Spatial structure of the catch composition

A nMDS representation of the sampling data ([Fig. 2](#)) showed a fair fit to the distinct groupings corresponding to the five different depth ranges. The PERMANOVA test detected significant differences among depths ($p = 0.001$), zones ($p = 0.001$), and seasons ($p = 0.001$). The species composition of the catch was significantly different among all depths, as detailed in [Table 2A](#) ($p < 0.01$ in all cases). Within the same depths, there were significant differences between the three zones ([Table 2B](#)). In particular, the deep shelf was different for each zone (north, center and south, $p < 0.01$ in all cases). The upper slope and the lower slope had significant differences between the north and the center zones ($p < 0.01$ in both cases). There were also significant differences among seasons within a same depth ([Table 2C](#)).

The SIMPER analysis identified the commercial species that characterized the different depths ([Table 2](#)). For example, the upper slope was characterized by the presence of *P. blennoides*, *N. norvegicus*, and *P. longirostris*, and the lower slope by *A. antennatus* ([Table 2A](#)). In

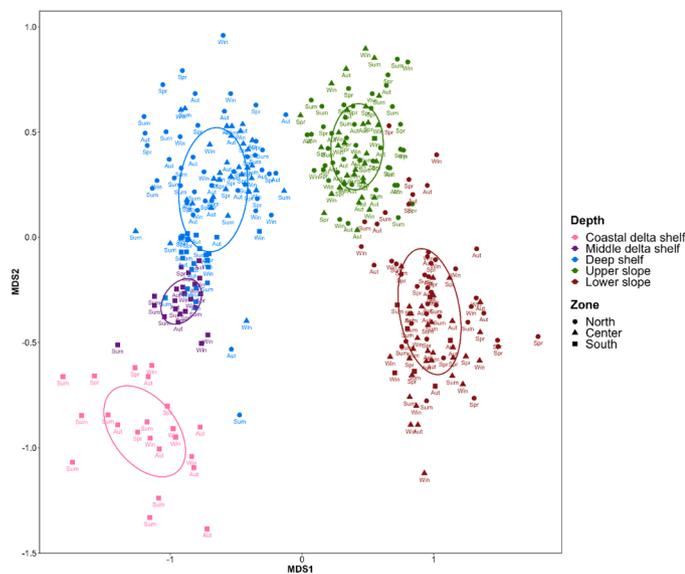


Fig. 2. Multidimensional ordination plot (nMDS; stress = 0.1517) of the single hauls. Colors represent depths (pink; coastal delta shelf, purple; middle delta shelf, blue; deep shelf, green; upper slope and red; lower slope); shapes represent zones (circles; north, triangles; center and squares; south), and letters represent seasons (Win; winter; Spr; spring, Aut; autumn and Sum; summer).

contrast, the shallower zones showed some overlapping, with the presence of the genus *Trachurus* as common species throughout the three ranges but with significant differences for other species (Table 2A). When the community composition was analyzed by depth and zone, the presence of *P. longirostris* and *M. surmuletus* characterized the landed fraction of the catch in the deep shelf of the north zone (Table 2B), while the dissimilarity in the northern samples is marked by the presence of a greater number of species. The upper and lower slopes of the center zone were characterized by the presence of *M. merluccius* and *G. melastomus*, respectively (Table 2B). As for the comparison by season, significant differences in species composition were observed at some depths, e.g., the presence of *M. merluccius* explaining a relevant part of the differences in the upper and lower slopes (Supplementary Table 4).

3.2. Length frequency comparison

The annual length-frequency distributions of the seven targeted species showed significant differences throughout the studied years for *E. cirrhosa* and *M. merluccius* (Fig. 3). In detail, between the years 2019 and 2020, *E. cirrhosa* showed a higher abundance of individuals under 200 g in 2020 (Kolmogorov-Smirnoff $p = 0.04$), and *M. merluccius* showed a higher abundance of individuals below the Minimum Conservation Reference Size (20 cm) in 2020 (Kolmogorov-Smirnoff $p = 0.01$).

The comparison of length frequency distribution across sampling depths showed significant differences in the case of *M. barbatius*, the coastal delta shelf presents a length frequency distribution that was significantly more centered around MCRS and L50 values than in other depths (Fig. 4). For *M. merluccius*, the upper and lower slope were significantly different from the rest of depth ranges, with a clear predominance of individuals above the MCRS and L50 (Fig. 4). In the case of *N. norvegicus*, individuals were significantly more evenly distributed across sizes in the lower slope, and the species showed a distribution with a median around 30 mm cephalothorax length (CL) in the upper slope (Supplementary Fig. 4). Finally, *P. longirostris* showed a distribution with values centered above the MCRS and L50 values in the upper slope, significantly different than the rest of depth ranges (Supplementary Fig. 4).

The pairwise size comparisons by depth for *M. merluccius* and

M. barbatius are illustrated in Fig. 3. The overall trends and by factors, with detailed statistics and p -values for all tests is available in Supplementary Table 2. The comparison among seasons revealed significant differences in the length distribution data especially for *E. cirrhosa*, where all seasons were different except for autumn and winter, and *P. longirostris*, which seemed to present two types of distribution, one for spring and summer, with a predominance of larger individuals, and another for autumn and winter, with more individuals under the MCRS (Supplementary Table 2). Distributions were fairly homogenous throughout all seasons for *M. merluccius* and *N. norvegicus*, which only presented significant differences between spring and autumn (Supplementary Table 2). Regarding the analysis by zone, there were significant differences for *M. barbatius*, *E. cirrhosa*, *P. longirostris*, and *A. antennatus* for at least one pair of depths compared (Supplementary Table 2). All species had significant differences for at least one of the season combinations studied.

3.3. Abundance and biomass

In the comparison among the three studied years, three species, *M. merluccius*, *P. longirostris*, and *A. antennatus*, showed significant differences through time, the two former for both abundance and biomass values, and the latter only for biomass values (Fig. 5, Supplementary Table 3). No significant differences (ANOVA $p > 0.05$; Supplementary Table 3) were found for the other studied species over time. All species showed significant differences in abundance and biomass values by depth strata, and among them, *M. barbatius* *N. norvegicus* did not show differences in abundance for any other factor (Supplementary Table 3, Annex II).

As for the analysis by season, abundance and biomass values were fairly stable across the data. Abundance of *S. mantis* in autumn ($1041.89 \pm 232.61 \text{ ind}\cdot\text{km}^{-2}$) was significantly higher than in the spring ($670.74 \pm 187.36 \text{ ind}\cdot\text{km}^{-2}$) and lower than in the summer ($1744.91 \pm 313.99 \text{ ind}\cdot\text{km}^{-2}$; Fig. 6, see Supplementary Table 3 for p -values). As for biomass, *E. cirrhosa* showed significantly higher values in the winter ($15.98 \pm 2.69 \text{ kg}\cdot\text{km}^{-2}$) than in the autumn ($12.54 \pm 2.34 \text{ kg}\cdot\text{km}^{-2}$) and in the summer ($6.90 \pm 1.19 \text{ kg}\cdot\text{km}^{-2}$). For *A. antennatus*, biomass was significantly higher in the summer ($47.18 \pm 9.72 \text{ kg}\cdot\text{km}^{-2}$) than in the autumn ($27.21 \pm 4.97 \text{ kg}\cdot\text{km}^{-2}$; Fig. 6; Supplementary Table 3).

Comparison among depth strata showed the most significant differences in all species. For *E. cirrhosa*, abundance and biomass values were significantly higher in the deep shelf ($100 \text{ ind}\cdot\text{km}^{-2}$ and $20 \text{ kg}\cdot\text{km}^{-2}$) and lower in the coastal delta shelf ($5 \text{ ind}\cdot\text{km}^{-2}$ and $1 \text{ kg}\cdot\text{km}^{-2}$; Fig. 6). In the case of *M. merluccius*, abundance and biomass showed significantly higher values in the deep shelf ($1191 \text{ ind}\cdot\text{km}^{-2}$ and $58 \text{ kg}\cdot\text{km}^{-2}$) than in all other strata except for the middle delta shelf. For *N. norvegicus* and *A. antennatus*, abundance and biomass showed an opposite trend, being significantly higher in the upper slope for the former ($2184.58 \pm 220.58 \text{ ind}\cdot\text{km}^{-2}$ and $49.26 \pm 4.84 \text{ kg}\cdot\text{km}^{-2}$ in average) and in the lower slope for the latter ($4120.03 \pm 268.52 \text{ ind}\cdot\text{km}^{-2}$ and $74.78 \pm 4.94 \text{ kg}\cdot\text{km}^{-2}$, ANOVA $p < 0.01$ in all cases).

4. Discussion

This study analyzes the spatial and temporal patterns affecting the key fishing resources exploited by the Catalan bottom trawling fleet, using data obtained from a newly implemented monitoring program along the Catalan coast. The dataset resulting from the monitoring program has proven to be generally comparable throughout the years and clearly reflects the variability through the seasons, zones and depth ranges for the targeted species. For instance, the commercial fraction of the catch of the continental slope shows predominance of a single species in the Center zone (*M. merluccius* in the upper slope, and *G. melastomus* in the lower slope), whereas in the North zone, the differences are explained by a group of species, and distinguished from the center by the presence of *N. norvegicus* in the upper slope and

Table 2

List of species contributing to the dissimilarity in the composition of the landed catch among levels of A: depth; and B: zone, obtained from SIMPER test. First two columns show the two levels of the factor that are significantly different. Spp 1 and 2 refer to the species that are most abundant in each level of reference. Dissimilarity refers to the total contribution of the species shown (in cumulative percentage) to the difference between levels 1 and 2 of each factor. p-value indicates the results of PERMANOVA test comparing each pair of levels.

| A | | | | | | |
|-------------------------|--------------------|---|---|---------------|-----------|--|
| Depths compared | | Species accounting for differences | | | PERMANOVA | |
| Depth 1 | Depth 2 | Spp 1 | Spp 2 | Dissimilarity | p-value | |
| Coastal delta shelf | Middle delta shelf | <i>Pagellus erythrinus</i> <i>Trachurus mediterraneus</i> <i>Squilla mantis</i> <i>Sphyræna sphyræna</i> <i>Mullus barbatus</i> | <i>Merluccius merluccius</i> <i>Illex coindetii</i> <i>Trachurus trachurus</i> <i>Eledone cirrhosa</i> | 46 % | < 0.01 | |
| Coastal delta shelf | Deep shelf | <i>Pagellus erythrinus</i> <i>Trachurus mediterraneus</i> <i>Squilla mantis</i> | <i>Trachurus trachurus</i> <i>Lophius budegassa</i> <i>Merluccius merluccius</i> <i>Mullus barbatus</i> | 47 % | < 0.01 | |
| Coastal delta shelf | Upper slope | <i>Pagellus erythrinus</i> <i>Squilla mantis</i> <i>Trachurus mediterraneus</i> | <i>Phycis blennoides</i> <i>Nephrops norvegicus</i> <i>Micromesistius poutassou</i> <i>Parapenaeus longirostris</i> | 48 % | < 0.01 | |
| Coastal delta shelf | Lower slope | <i>Pagellus erythrinus</i> <i>Squilla mantis</i> <i>Trachurus mediterraneus</i> | <i>Aristeus antennatus</i> <i>Phycis blennoides</i> | 46 % | < 0.01 | |
| Middle delta shelf | Deep shelf | <i>Squilla mantis</i> <i>Eledone cirrhosa</i> | <i>Trachurus trachurus</i> <i>Lophius budegassa</i> <i>Merluccius merluccius</i> <i>Mullus barbatus</i> <i>Illex coindetii</i> | 49 % | < 0.01 | |
| Middle delta shelf | Upper slope | <i>Merluccius merluccius</i> <i>Squilla mantis</i> <i>Eledone cirrhosa</i> | <i>Phycis blennoides</i> <i>Nephrops norvegicus</i> <i>Micromesistius poutassou</i> <i>Parapenaeus longirostris</i> | 47 % | < 0.01 | |
| Middle delta shelf | Lower slope | <i>Merluccius merluccius</i> <i>Squilla mantis</i> <i>Illex coindetii</i> <i>Eledone cirrhosa</i> | <i>Aristeus antennatus</i> <i>Phycis blennoides</i> | 46 % | < 0.01 | |
| Deep shelf | Upper slope | <i>Trachurus trachurus</i> <i>Lophius budegassa</i> <i>Mullus barbatus</i> <i>Merluccius merluccius</i> | <i>Phycis blennoides</i> <i>Nephrops norvegicus</i> <i>Parapenaeus longirostris</i> | 55 % | < 0.01 | |
| Deep shelf | Lower slope | <i>Trachurus trachurus</i> <i>Merluccius merluccius</i> <i>Lophius budegassa</i> <i>Mullus barbatus</i> | <i>Aristeus antennatus</i> | 50 % | < 0.01 | |
| Upper slope | Lower slope | <i>Nephrops norvegicus</i> <i>Phycis blennoides</i> <i>Micromesistius poutassou</i> <i>Parapenaeus longirostris</i> | <i>Aristeus antennatus</i> | 60 % | < 0.01 | |
| B | | | | | | |
| Depths & zones compared | | Species accounting for differences | | | PERMANOVA | |
| Depth & zone 1 | Depth & zone 2 | Spp 1 | Spp 2 | Dissimilarity | p-value | |
| Deep shelf North | Deep shelf Center | <i>Mullus barbatus</i> <i>Parapenaeus longirostris</i> <i>Trisopterus capelanus</i> | <i>Trachurus trachurus</i> <i>Lophius budegassa</i> <i>Illex coindetii</i> <i>Merluccius merluccius</i> <i>Eledone cirrhosa</i> | 66 % | 0.02 | |
| Deep shelf North | Deep shelf South | <i>Trachurus trachurus</i> <i>Mullus barbatus</i> <i>Illex coindetii</i> <i>Parapenaeus longirostris</i> | <i>Lophius budegassa</i> <i>Merluccius merluccius</i> <i>Eledone cirrhosa</i> | 52 % | <0.01 | |
| Deep shelf Center | Deep shelf South | <i>Trachurus trachurus</i> <i>Merluccius merluccius</i> <i>Illex coindetii</i> <i>Mullus barbatus</i> | <i>Lophius budegassa</i> <i>Eledone cirrhosa</i> | 54 % | <0.01 | |
| Upper slope North | Upper slope Center | <i>Phycis blennoides</i> <i>Nephrops norvegicus</i> <i>Micromesistius poutassou</i> <i>Parapenaeus longirostris</i> | <i>Merluccius merluccius</i> | 59 % | <0.01 | |
| Lower slope North | Lower slope Center | <i>Aristeus antennatus</i> <i>Phycis blennoides</i> <i>Merluccius merluccius</i> <i>Micromesistius poutassou</i> | <i>Galeus melastomus</i> | 66 % | <0.01 | |

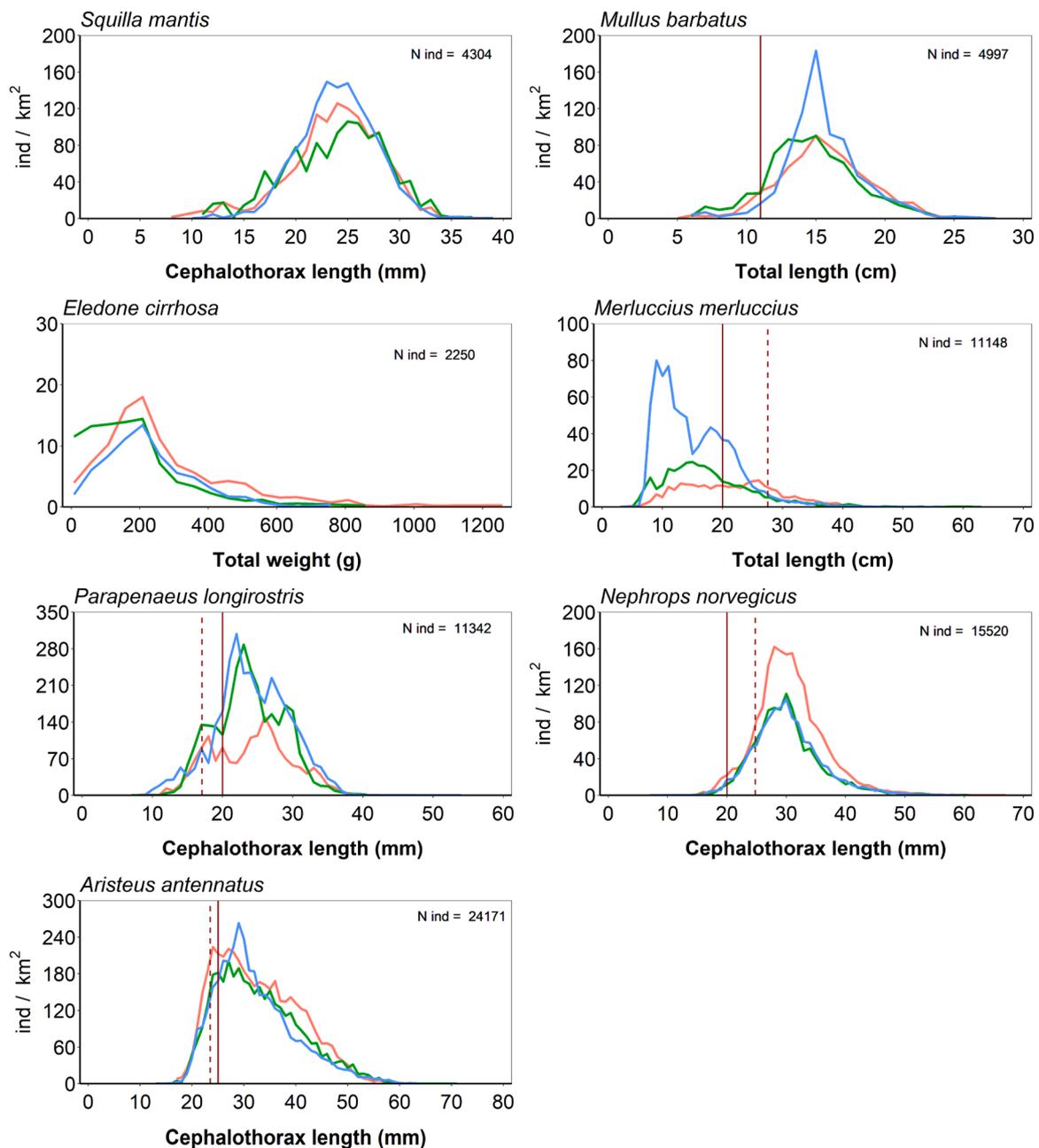


Fig. 3. Annual length-frequency distribution of seven target species for the years 2019 (red), 2020 (green), and 2021 (blue). Red line: Minimum Conservation Reference Size (MCRS). Red dashed line: size at first maturity obtained with ICATMAR data.

A. antennatus in the lower slope. The high abundance and biomass values for *A. antennatus* in the South zone correspond to the samplings in the lower slope, which is the depth of preference for this species. These results are of main importance because the collection of extensive and reliable data is the basis for fisheries assessment and the resulting datasets need to be up to scale with the particularities of each region [25, 26]. The design of a fisheries monitoring program should aim to collect data that reflect dynamics of both biological and spatio-temporal trends in different spatial extents and scales [27], as well as the seasonality that defines the changing fleet strategies throughout the year [28]. The locally-based monitoring program designed by ICATMAR and its data collection are capturing the particularities of the area, where fisheries are markedly multispecific and fleet behavior differs by area, being representative of local and regional characteristics that would be undetectable at a larger spatio-temporal scale monitoring.

The present sampling provides data on target species that are needed

for stock assessment analysis, such as abundance, length frequency distributions or size at first maturity. The published information on these parameters in the Mediterranean Sea is scarce and often comes from one-time studies, and as such, does not support seasonal or inter-annual comparison (e.g. [29–31]). For example, the significant increase of *M. merluccius* individuals below L50 in 2020 found in this study gives a glimpse of the interannual variability of the recruitment of this species, which can be favored by environmental factors such as enhanced primary production or late winter low temperatures promoting egg and larval survival [32]. The parameters may also vary according to locations. For example, a study from the different GSA areas in the Mediterranean Sea determined that Sardinia (GSA 11) had the highest abundance and biomass for *M. merluccius* with 180,493 ind·km⁻² and 4636 kg·km⁻² [33], values much above the ones found in this study. These type of data may not only vary with environmental changes, including climate change [34,35], but also with high fishing pressure,

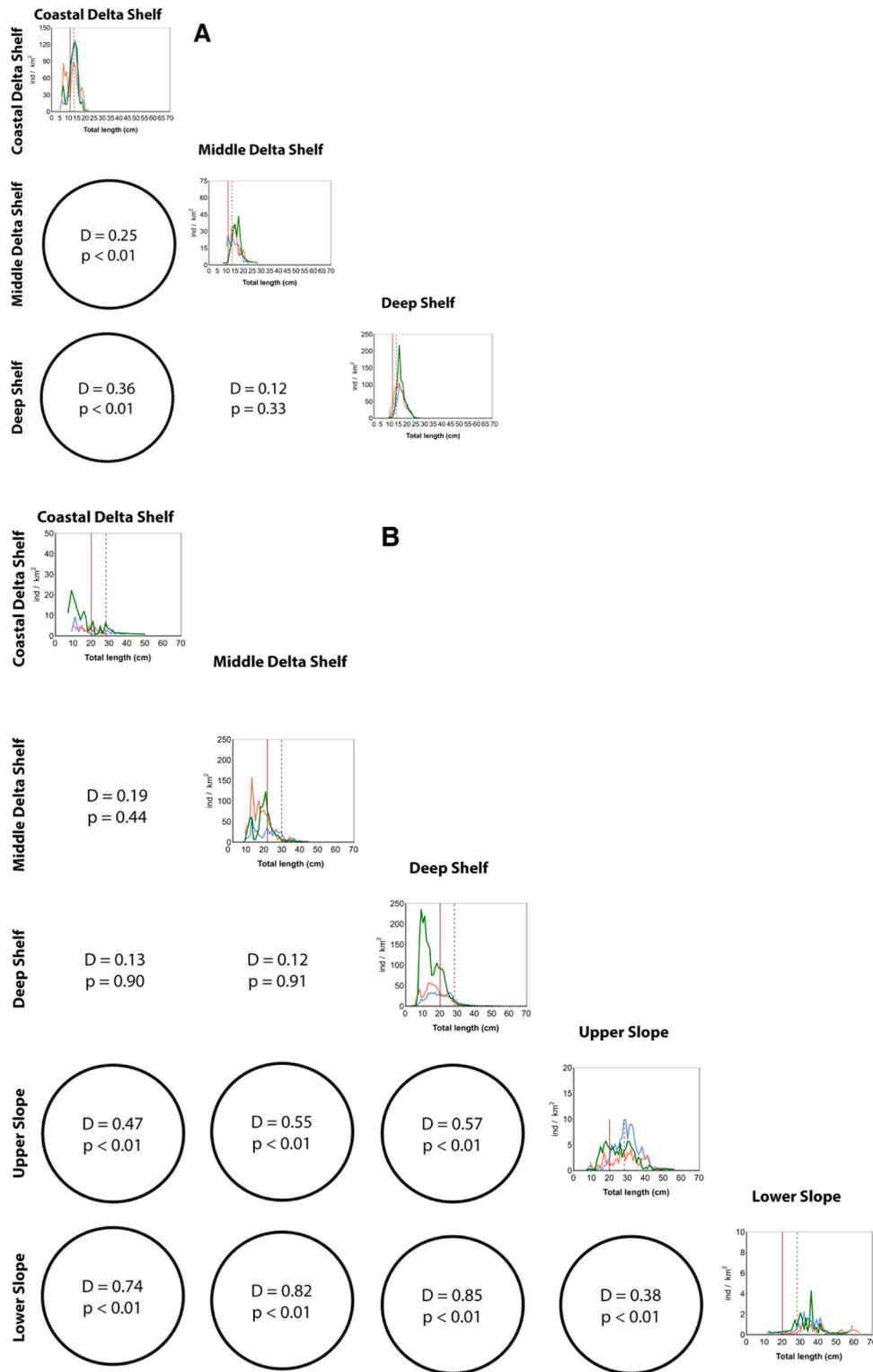


Fig. 4. Length-frequency distribution by depth for the years 2019–2021 for (A) *M. barbatus* and (B) *M. merluccius*. Red line: Minimum Conservation Reference Size (MCRS). Red dashed line: size at first maturity (L50) obtained with ICATMAR data. D: Kolmogorov’s D statistic; p: p-value of the pairwise test. Circles indicate significant differences between a pair of depths.

which may entail a decrease in the size at first maturity (L50) of over-exploited species [36,37]. For instance, the L50 for *N. norvegicus* inhabiting along the Catalan coast has been reduced in 5 mm compared to 1974 [37]. On the contrary, *M. barbatus* does not seem to have changed over time since a study in the Mediterranean (1994–2000) determined that most catches were composed of animals smaller than 15 cm [38] and the L50 has been determined within the range of 13 cm

(for females) in different studies from different time periods [39–41].

The sampling on board bottom trawling vessels provides length data across the depth range for the studied species. However, some gaps on species population characteristics might be difficult to bridge when only using bottom trawling. This is the case of larger individuals of *M. merluccius*, which are missing from the sampling because they are caught by small-scale fisheries (longline and trammel nets) [42]. Also,

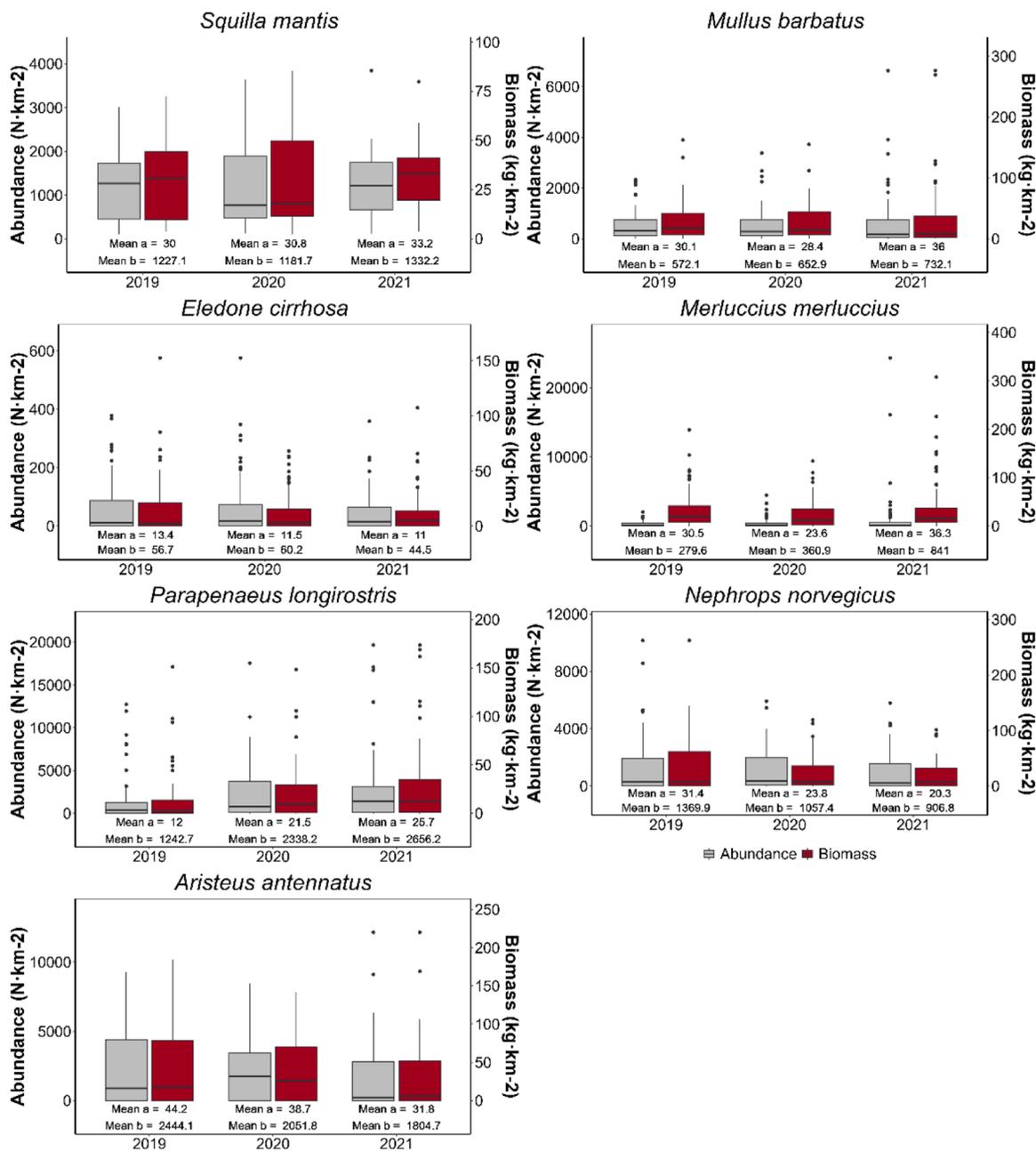


Fig. 5. Box plots showing the abundance (ind·km⁻², in gray) and biomass (kg·km⁻², in red) of the seven targeted species for the studied years (2019–2021). Mean a: mean abundance; Mean b: mean biomass. Regarding the different zones, biomass of *E. cirrhosa* was significantly lower in the center zone (9.49 ± 1.41 kg·km⁻²) than in the north and south zones (11.60 ± 1.90 and 16.11 ± 2.20 kg·km⁻² respectively; Fig. 6; ANOVA $p < 0.01$). Abundance of *A. antennatus* was significantly lower in the center zone (1624.83 ± 252.98 ind·km⁻²; Fig. 6) than in the other zones, reaching a maximum in the south zone with 4515.84 ± 919.96 ind·km⁻² (Fig. 6).

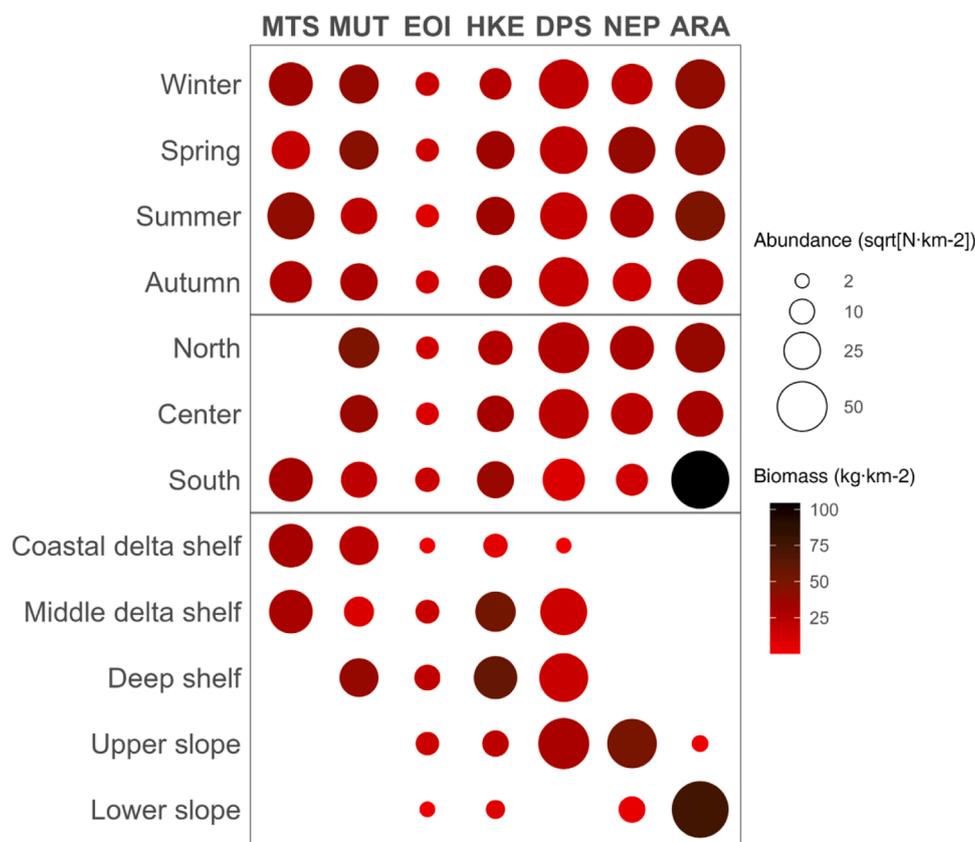


Fig. 6. Average abundance (in $\sqrt{[N \cdot \text{km}^{-2}]}$) and biomass (in $\text{kg} \cdot \text{km}^{-2}$) for the seven studied species attending to factors season (top), zone (center), and depth (bottom). Size of the bubbles indicates abundance, and color indicates biomass. MTS: *S. mantis*; MUT: *M. barbatus*; EOI: *E. cirrhosa*; HKE: *M. merluccius*; DPS: *P. longirostris*; NEP: *N. norvegicus*; ARA: *A. antennatus*.

the depth range of *M. barbatus* includes shallower waters than those sampled in this study [43], and changes are already underway in the sampling structure to improve the coverage of this species, along with a plan to integrate small-scale fisheries in the monitoring program. In all, the monitoring program obtains biological information about where most of the species are distributed and provides proof that the fishing resources do not have a homogenous distribution in the Catalan margin. The potential of studying the relationship between fishing resources distribution and environmental variables using these continuous monitoring data has been laid out in the present study, but needs to be further explored in subsequent works that can address each species separately and in depth.

Besides information on target species, the monitoring program provides other key data for a better governance of the area including information on other commercial but non-target species, discards, marine litter or the identification of new species and resources. A recent study using ICATMAR data found an average discard ratio of 25 % for bottom trawling in the northern GSA 6 [8]. Discards assessment, considering both undersized individuals of commercial species and non-commercial species, is essential to evaluate the impact of fisheries in the ecosystems and implement ecosystem-based fisheries management actions [44]. In the framework of the Marine Strategy Framework Directive (MSFD) of the European Commission to achieve Good Environmental Status (GES) [46], an essential step is to establish monitoring programs for assessment enabling the state of the marine waters to be evaluated on a regular basis [47].

The current management strategies in the Mediterranean Sea derive from a North Atlantic fishery model, which are revised every year based on monitoring strategies from the EU Data Collection Framework (i.e., data from DCF MEDITS experimental cruises). However, regulations that are effective in the North Atlantic may not fit other fisheries in the

European Union [45]. The Mediterranean Sea marine habitats are different than those from other European waters and may be defined by specific ecological characteristics [46]. Moreover, the Mediterranean has a typical seasonal pattern of temperate regions, with a thermocline from May to November and vertically mixed waters the rest of the year [47]. All these characteristics favor a spatial and temporal distribution of the marine resources, including key species for the bottom trawl fisheries [33,48,49]. Moreover, the northern GSA 6 is not only distinct by the distribution of its fishery resources, but also by its social component. The area is organized by fishery guilds, which were defined in the Middle Ages, and currently play a social and economic function in managing the resources [50]. These guilds can, as they have done in the past, implement local management strategies such as requiring a larger mesh size (50-mm instead of 40-mm square) for the blue and red shrimp fishery in Palamós (BOE 2018, APM/532/2018). The effectivity of local management strategies is only quantifiable with detailed long-term monitoring programs such as ICATMAR's, which began in 2019 and aims to continuously monitor the resources from the Catalan margin in the long term.

Managing the fishing resources of the Mediterranean is not an easy task. The species and fishing patterns are quite diverse, with landing sites widely distributed geographically [45] along a coastal region where fishing activity is one of the main socioeconomic activities in many small-medium localities. Although management strategies have been implemented for decades, their outcomes have not been as successful as expected for different reasons, including the disregard for scientific advice and the deficiencies of current national management plans [25]. Exhaustive locally-based scientific monitoring programs may be key to fisheries sustainability in the Mediterranean basin, especially when structured within more regionalized management areas that can accurately detect the changing patterns of fishing resources.

5. Conclusion

The main fishing resources exploited by the Catalan bottom trawling fleet, i.e. *S. mantis*, *M. barbatus*, *E. cirrhosa*, *M. merluccius*, *P. longirostris*, *N. norvegicus*, and *A. antennatus*, are not homogeneously distributed in time or space throughout the northern GSA 6. An exhaustive continuous monitoring program allowed for higher spatial and temporal resolution of biological data on target species reflecting the dynamics of the fleet. We suggest that this type of sampling be included in Mediterranean areas to complement the current DCF program, so that fisheries can be managed evaluating the biology of the species along with the social component of the fisheries structure.

CRediT authorship contribution statement

Antoni Lombarte: Writing – review & editing, Conceptualization. **Alberto J Rico:** Writing – review & editing, Investigation. **Mariona Garriga-Panisello:** Writing – review & editing, Formal analysis. **Jordi Ribera-Altimir:** Writing – review & editing, Software. **Marta Blanco:** Writing – review & editing, Investigation, Formal analysis. **Joan Sala-Coromina:** Writing – review & editing, Visualization. **Marc Balcells:** Writing – review & editing, Investigation. **Alba Rojas:** Writing – review & editing, Investigation. **Ana I. Colmenero:** Writing – review & editing,

Annex I. – Laboratory protocols by species group

Crustaceans

The following data is taken from each of the 30 individuals of blue and red shrimp, deep-water rose shrimp, Norway lobster and spottail mantis shrimp fished and transported to the laboratory: cephalothorax length (mm), total weight (0.01 g), sex (male, female, or indeterminate), gonad weight (0.01 g) and sexual maturity for females. For the spottail mantis shrimp, sexual maturity stages are 1: immature/resting, 2: maturing, and 3: mature. For the other target crustaceans, the stages are defined as 1: immature, 2: resting, 3: maturing, 4: mature, 5: spawning. Other specific parameters measured include the presence of a spermatophore for the blue and red shrimp, the stage of the eggs for Norway lobster (1: just laid, 2: developing, 3: developed) and the stage of the female cement glands for the spottail mantis shrimp (1: immature/resting, 2: maturing, 3: mature).

Horned octopus

A sample of 30 individuals are taken for biological sampling. The measurements taken are dorsal mantle length (mm), total weight (0.1 g), sex (male or female), and sexual maturity (1: immature, 2: developing, 3: maturing, 4: mature, 5: spawning; 6: post spawning). For females, the presence of a spermatangium in the ovary is recorded.

European hake and red mullet

All individuals caught are measured (mm), 30 individuals for each category size are taken for biological sampling recording the following data: total length (mm), total weight (0.1 g), gutted weight (0.1 g), sex (male, female, or indeterminate), sexual maturity (1: immature, 2: resting, 3: maturing, 4: advanced maturation, 5: spawning; 6: post spawning), gonad weight (0.01 g), and, for the hake, liver weight (0.01 g), and stomach state (empty, full, evaginated empty, or evaginated full).

Discard fraction

The discarded fraction is preserved at -20°C until further analysis, when all individuals are identified to the lowest possible taxonomic level, and individually measured and weighed using at least one decimal place (± 0.1 g), in all cases where possible. Each individual is measured (mm) following these criteria:

- fish: total length (anal length for macrurids)
- crustaceans: cephalothorax length
- cephalopods: dorsal mantle length (length to mid-eye for octopus)
- bivalves *Acanthocardia* spp. and *Venus nux*: total length
- gastropods: total length for *Galeodea* spp., length without siphon *Bolinus brandaris*

The natural debris, which include organic materials from both terrestrial and marine origin, are weighted (± 0.1 g) according to seven different categories: calcareous debris, marine algae, marine organic, marine plants, shells, terrestrial animals, and terrestrial plants. Marine litter, i.e. items with an anthropogenic origin either whole or broken, is classified following the application of Directive EU 2019/883 of the European Parliament and of the Council as regards to monitoring data methodologies and the format for reporting passively fished waste (EU 2022/92).

Project administration. **Joan B. Company:** Writing – review & editing, Supervision, Funding acquisition, Conceptualization. **Marta Pujol-Baucells:** Writing – review & editing, Investigation. **Laura Recasens:** Writing – review & editing, Funding acquisition, Conceptualization. **Xènia Puigcerver-Segarra:** Writing – review & editing, Investigation. **Mireia G. Mingote:** Writing – review & editing, Investigation, Formal analysis. **Eve Galimany:** Writing – original draft, Investigation, Conceptualization. **John G. Ramírez:** Writing – review & editing, Conceptualization. **Marta Carreton:** Writing – original draft, Formal analysis, Conceptualization. **Mireia Silvestre:** Writing – review & editing, Investigation. **David Nos:** Writing – review & editing, Investigation. **Ricardo Santos-Bethencourt:** Writing – review & editing, Investigation. **Ferran Bustos:** Writing – review & editing, Investigation. **Cristina López-Pérez:** Writing – review & editing, Investigation.

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Annex II. – Information on the implementation of the General Linear Models

Model formulas in all cases are:

glm.nb(abundancia ~ Any + Estacio + ZonaPort + TipusDeFons, data = data.analysis)

glm.nb(biomassa ~ Any + Estacio + ZonaPort + TipusDeFons, data = data.analysis)

Information presented for each species:

- Quantile-Quantile (Q-Q) plot of the residuals
- Summary of the model
- Summary of the ANOVA analysis of the model

Any = year

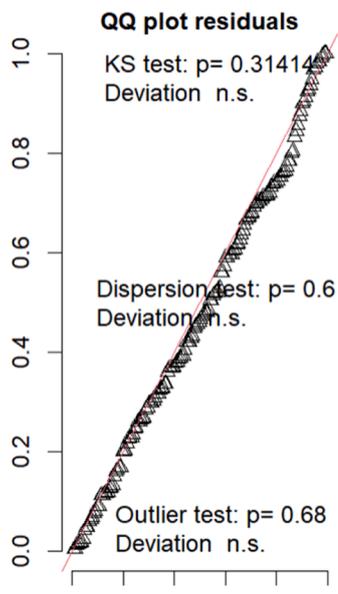
Estacio = season

ZonaPort = zone

TipusDeFons = depth

MTS (*Squilla mantis*)

MTS – Abundance



Deviance Residuals:

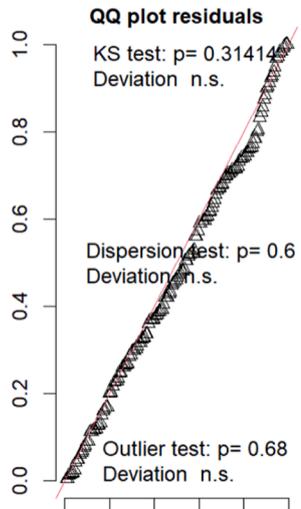
| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -2.8733 | -0.6669 | -0.5399 | 0.1203 | 4.3948 |

Coefficients:

| | Estimate | std. Error | z value | Pr(> z) | |
|------------------------|----------|------------|---------|----------|-----|
| (Intercept) | 3.87104 | 0.15197 | 25.473 | < 2e-16 | *** |
| Any2020 | -0.01914 | 0.14148 | -0.135 | 0.89239 | |
| Any2021 | -0.27715 | 0.14182 | -1.954 | 0.05067 | . |
| EstacioSpring | 0.42935 | 0.16238 | 2.644 | 0.00819 | ** |
| EstacioSummer | 0.46131 | 0.15716 | 2.935 | 0.00333 | ** |
| EstacioWinter | 0.36378 | 0.16679 | 2.181 | 0.02918 | * |
| ZonaPortNorth | 0.27225 | 0.12162 | 2.239 | 0.02519 | * |
| ZonaPortSouth | 0.63203 | 0.22632 | 2.793 | 0.00523 | ** |
| TipusDeFonsUpper slope | -5.96102 | 0.27504 | -21.674 | < 2e-16 | *** |

| | Df | Deviance | Resid. Df | Resid. Dev | F | Pr(>F) | |
|-------------|----|----------|-----------|------------|-----------|-----------|-----|
| NULL | | | 161 | 1635.72 | | | |
| Any | 2 | 10.26 | 159 | 1625.46 | 5.1295 | 0.005919 | ** |
| Estacio | 3 | 23.25 | 156 | 1602.21 | 7.7492 | 3.586e-05 | *** |
| ZonaPort | 2 | 36.41 | 154 | 1565.80 | 18.2070 | 1.238e-08 | *** |
| TipusDeFons | 1 | 1417.62 | 153 | 148.18 | 1417.6181 | < 2.2e-16 | *** |

MTS – Biomass



Deviance Residuals:

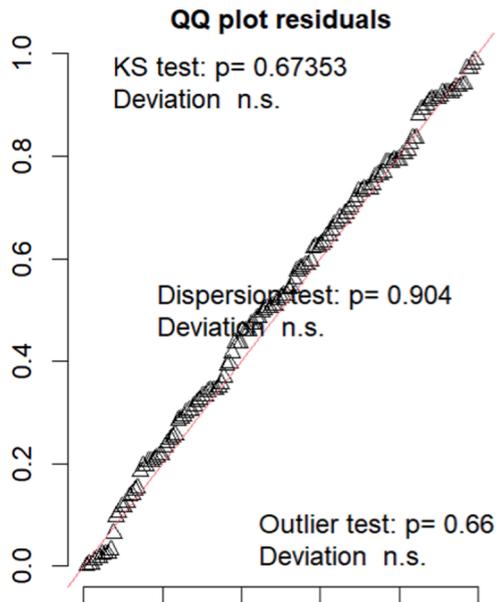
| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -2.8733 | -0.6669 | -0.5399 | 0.1203 | 4.3948 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) |
|------------------------|----------|------------|---------|-------------|
| (Intercept) | 3.87104 | 0.15197 | 25.473 | < 2e-16 *** |
| Any2020 | -0.01914 | 0.14148 | -0.135 | 0.89239 |
| Any2021 | -0.27715 | 0.14182 | -1.954 | 0.05067 . |
| EstacioSpring | 0.42935 | 0.16238 | 2.644 | 0.00819 ** |
| EstacioSummer | 0.46131 | 0.15716 | 2.935 | 0.00333 ** |
| EstacioWinter | 0.36378 | 0.16679 | 2.181 | 0.02918 * |
| ZonaPortNorth | 0.27225 | 0.12162 | 2.239 | 0.02519 * |
| ZonaPortSouth | 0.63203 | 0.22632 | 2.793 | 0.00523 ** |
| TipusDeFonsUpper slope | -5.96102 | 0.27504 | -21.674 | < 2e-16 *** |

| | Df | Deviance | Resid. Df | Resid. Dev | F | Pr(>F) |
|-------------|----|----------|-----------|------------|-----------|---------------|
| NULL | | | 161 | 1635.72 | | |
| Any | 2 | 10.26 | 159 | 1625.46 | 5.1295 | 0.005919 ** |
| Estacio | 3 | 23.25 | 156 | 1602.21 | 7.7492 | 3.586e-05 *** |
| ZonaPort | 2 | 36.41 | 154 | 1565.80 | 18.2070 | 1.238e-08 *** |
| TipusDeFons | 1 | 1417.62 | 153 | 148.18 | 1417.6181 | < 2.2e-16 *** |

MUT (*Mullus barbatus*)
 MUT – Abundance



Coefficients:

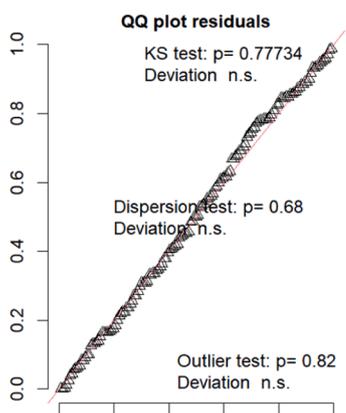
| | Estimate | Std. Error | z value | Pr(> z) |
|--------------------------------------|----------|------------|---------|-------------|
| (Intercept) | 6.33866 | 0.32397 | 19.566 | < 2e-16 *** |
| Any2020 | 0.17623 | 0.26952 | 0.654 | 0.51321 |
| Any2021 | 0.22046 | 0.25743 | 0.856 | 0.39179 |
| EstacioSpring | 0.04430 | 0.31019 | 0.143 | 0.88644 |
| EstacioSummer | -0.35967 | 0.29100 | -1.236 | 0.21646 |
| EstacioWinter | 0.03199 | 0.31141 | 0.103 | 0.91819 |
| ZonaPortNorth | 0.32581 | 0.29851 | 1.091 | 0.27506 |
| ZonaPortSouth | -0.06013 | 0.35409 | -0.170 | 0.86515 |
| TipusDeFonsMiddle continental shelf | -1.00063 | 0.38792 | -2.579 | 0.00989 ** |
| TipusDeFonsShallow continental shelf | 0.46397 | 0.38769 | 1.197 | 0.23140 |

 Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

| | Df | Deviance | Resid. Df | Resid. Dev | F | Pr(>F) |
|-------------|----|----------|-----------|------------|--------|-------------|
| NULL | | | 144 | 196.69 | | |
| Any | 2 | 0.9370 | 142 | 195.76 | 0.4685 | 0.625946 |
| Estacio | 3 | 2.0951 | 139 | 193.66 | 0.6984 | 0.552902 |
| ZonaPort | 2 | 3.0748 | 137 | 190.59 | 1.5374 | 0.214939 |
| TipusDeFons | 2 | 11.9065 | 135 | 178.68 | 5.9533 | 0.002597 ** |

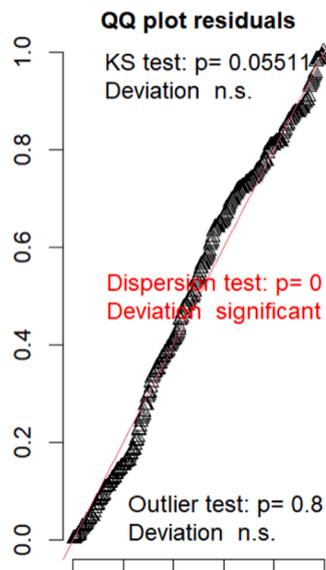
 Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

MUT – Biomass



| | Df | Deviance | Resid. | Df | Resid. Dev | F | Pr(>F) |
|---|----|----------|--------|-----|------------|--------|------------|
| NULL | | | | 144 | 202.27 | | |
| Any | 2 | 1.1031 | | 142 | 201.17 | 0.5516 | 0.57605 |
| Estacio | 3 | 9.7242 | | 139 | 191.45 | 3.2414 | 0.02106 * |
| ZonaPort | 2 | 11.0080 | | 137 | 180.44 | 5.5040 | 0.00407 ** |
| TipusDeFons | 2 | 9.0547 | | 135 | 171.38 | 4.5273 | 0.01081 * |
| --- | | | | | | | |
| Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 | | | | | | | |

EOI (*Eledone cirrhosa*)
EOI – Abundance



Deviance Residuals:

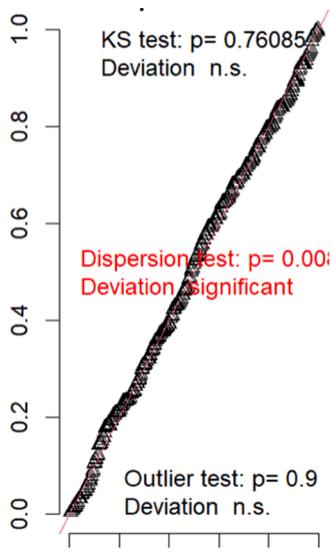
| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -1.9163 | -1.2590 | -0.5383 | 0.1371 | 2.1786 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) |
|--------------------------------------|----------|------------|---------|--------------|
| (Intercept) | 4.01198 | 0.33575 | 11.949 | < 2e-16 *** |
| Any2020 | 0.25906 | 0.26477 | 0.978 | 0.3279 |
| Any2021 | -0.17995 | 0.25586 | -0.703 | 0.4819 |
| EstacioSpring | 0.12468 | 0.30074 | 0.415 | 0.6785 |
| EstacioSummer | 0.03221 | 0.29000 | 0.111 | 0.9115 |
| EstacioWinter | 0.55827 | 0.30731 | 1.817 | 0.0693 . |
| ZonaPortNorth | 0.59198 | 0.24813 | 2.386 | 0.0170 * |
| ZonaPortSouth | 0.70011 | 0.39835 | 1.758 | 0.0788 . |
| TipusDeFonsLower slope | -3.05241 | 0.28275 | -10.796 | < 2e-16 *** |
| TipusDeFonsMiddle continental shelf | -0.71210 | 0.51928 | -1.371 | 0.1703 |
| TipusDeFonsShallow continental shelf | -3.59826 | 0.53082 | -6.779 | 1.21e-11 *** |
| TipusDeFonsUpper slope | -0.49707 | 0.28794 | -1.726 | 0.0843 . |

| | Df | Deviance | Resid. Df | Resid. Dev | F | Pr(>F) |
|-------------|----|----------|-----------|------------|---------|------------|
| NULL | | | 306 | 440.76 | | |
| Any | 2 | 1.467 | 304 | 439.29 | 0.7337 | 0.4801 |
| Estacio | 3 | 0.544 | 301 | 438.75 | 0.1813 | 0.9092 |
| ZonaPort | 2 | 2.800 | 299 | 435.95 | 1.4001 | 0.2466 |
| TipusDeFons | 4 | 107.831 | 295 | 328.12 | 26.9577 | <2e-16 *** |

EOI – Biomass



Deviance Residuals:

| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -1.9882 | -1.1476 | -0.6200 | 0.1678 | 2.2960 |

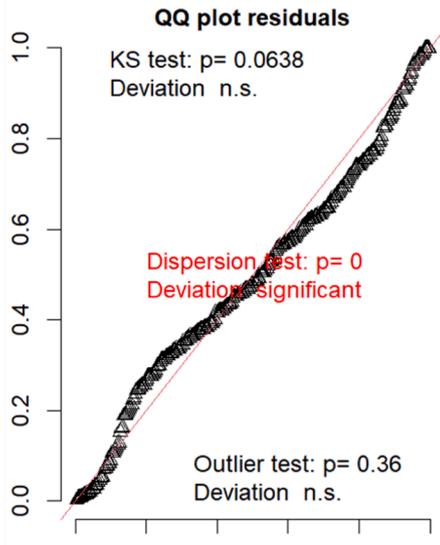
Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) | |
|--------------------------------------|----------|------------|---------|----------|-----|
| (Intercept) | 2.34101 | 0.28357 | 8.256 | < 2e-16 | *** |
| Any2020 | 0.02047 | 0.22594 | 0.091 | 0.92782 | |
| Any2021 | -0.19020 | 0.21853 | -0.870 | 0.38411 | |
| EstacioSpring | 0.18115 | 0.25581 | 0.708 | 0.47885 | |
| EstacioSummer | -0.40058 | 0.25082 | -1.597 | 0.11025 | |
| EstacioWinter | 0.71449 | 0.25960 | 2.752 | 0.00592 | ** |
| ZonaPortNorth | 0.59674 | 0.21191 | 2.816 | 0.00486 | ** |
| ZonaPortSouth | 1.02607 | 0.33302 | 3.081 | 0.00206 | ** |
| TipusDeFonsLower slope | -2.46035 | 0.24671 | -9.973 | < 2e-16 | *** |
| TipusDeFonsMiddle continental shelf | -0.53066 | 0.42802 | -1.240 | 0.21505 | |
| TipusDeFonsShallow continental shelf | -3.67572 | 0.49266 | -7.461 | 8.59e-14 | *** |
| TipusDeFonsUpper slope | -0.07642 | 0.23764 | -0.322 | 0.74778 | |

| | Df | Deviance | Resid. Dev | Df Resid. Dev | F | Pr(>F) |
|-------------|----|----------|------------|---------------|--------|-----------------------|
| NULL | | | | 306 | 475.69 | |
| Any | 2 | 1.042 | | 304 | 474.65 | 0.5211 0.593861 |
| Estacio | 3 | 12.915 | | 301 | 461.73 | 4.3051 0.004824 ** |
| ZonaPort | 2 | 9.396 | | 299 | 452.34 | 4.6980 0.009114 ** |
| TipusDeFons | 4 | 125.094 | | 295 | 327.24 | 31.2735 < 2.2e-16 *** |

HKE (*Merluccius merluccius*)

HKE - Abundance



Deviance Residuals:

| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -2.6884 | -0.9005 | -0.4472 | 0.0860 | 4.5245 |

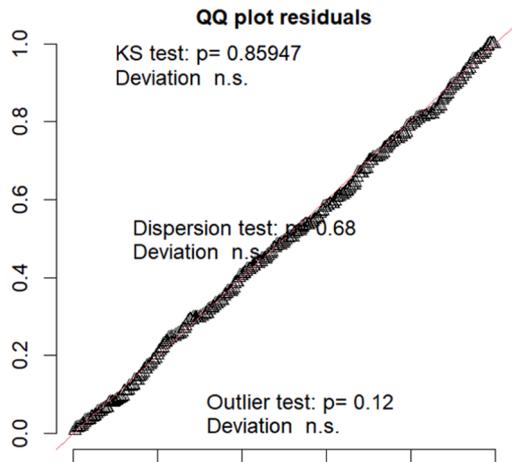
Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) | |
|--------------------------------------|----------|------------|---------|----------|-----|
| (Intercept) | 7.13280 | 0.21327 | 33.445 | < 2e-16 | *** |
| Any2020 | -0.29357 | 0.16847 | -1.743 | 0.08141 | . |
| Any2021 | 0.43452 | 0.16194 | 2.683 | 0.00729 | ** |
| EstacioSpring | 0.08828 | 0.19060 | 0.463 | 0.64324 | |
| EstacioSummer | 0.11856 | 0.18367 | 0.646 | 0.51857 | |
| EstacioWinter | -0.28002 | 0.19548 | -1.432 | 0.15200 | |
| ZonaPortNorth | -0.44619 | 0.15744 | -2.834 | 0.00460 | ** |
| ZonaPortSouth | -0.28780 | 0.25395 | -1.133 | 0.25710 | |
| TipusDeFonsLower slope | -4.04502 | 0.17927 | -22.564 | < 2e-16 | *** |
| TipusDeFonsMiddle continental shelf | -0.14167 | 0.33115 | -0.428 | 0.66880 | |
| TipusDeFonsShallow continental shelf | -2.74988 | 0.33221 | -8.277 | < 2e-16 | *** |
| TipusDeFonsUpper slope | -2.26945 | 0.18380 | -12.347 | < 2e-16 | *** |

| | Df | Deviance | Resid. Dev | Df Resid. Dev | F | Pr(>F) | |
|-------------|----|----------|------------|---------------|--------|----------|---------------|
| NULL | | | | 306 | 886.46 | | |
| Any | 2 | 53.69 | | 304 | 832.78 | 26.8437 | 2.198e-12 *** |
| Estacio | 3 | 17.29 | | 301 | 815.49 | 5.7623 | 0.0006169 *** |
| ZonaPort | 2 | 21.90 | | 299 | 793.59 | 10.9522 | 1.752e-05 *** |
| TipusDeFons | 4 | 420.57 | | 295 | 373.02 | 105.1423 | < 2.2e-16 *** |

 Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

HKE – Biomass



Deviance Residuals:

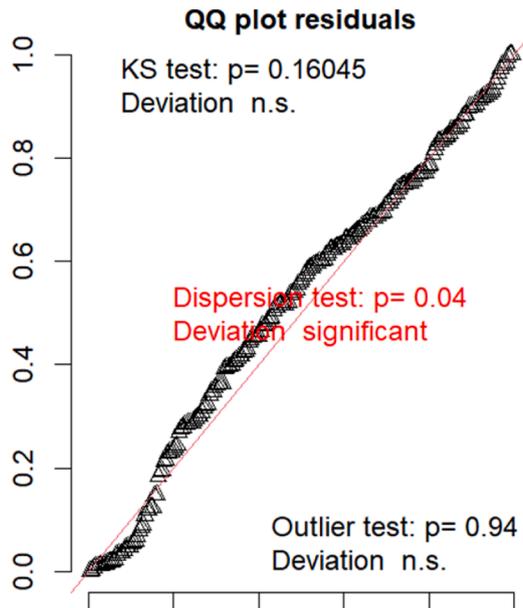
| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -2.4334 | -0.9599 | -0.3346 | 0.3526 | 4.1049 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) |
|--------------------------------------|----------|------------|---------|--------------|
| (Intercept) | 4.46473 | 0.17659 | 25.284 | < 2e-16 *** |
| Any2020 | -0.44871 | 0.14111 | -3.180 | 0.00147 ** |
| Any2021 | -0.04998 | 0.13486 | -0.371 | 0.71092 |
| EstacioSpring | -0.19372 | 0.15886 | -1.219 | 0.22267 |
| EstacioSummer | -0.10969 | 0.15310 | -0.716 | 0.47370 |
| EstacioWinter | -0.38082 | 0.16359 | -2.328 | 0.01991 * |
| ZonaPortNorth | -0.17407 | 0.13125 | -1.326 | 0.18476 |
| ZonaPortSouth | -0.18820 | 0.21059 | -0.894 | 0.37147 |
| TipusDeFonsLower slope | -2.11879 | 0.15053 | -14.076 | < 2e-16 *** |
| TipusDeFonsMiddle continental shelf | -0.03761 | 0.27227 | -0.138 | 0.89013 |
| TipusDeFonsShallow continental shelf | -2.27397 | 0.28569 | -7.959 | 1.73e-15 *** |
| TipusDeFonsUpper slope | -1.02408 | 0.15166 | -6.753 | 1.45e-11 *** |

| | Df | Deviance | Resid. Df | Resid. Dev | F | Pr(>F) |
|-------------|----|----------|-----------|------------|---------|---------------|
| NULL | | | 306 | 617.46 | | |
| Any | 2 | 9.361 | 304 | 608.10 | 4.6803 | 0.009276 ** |
| Estacio | 3 | 4.533 | 301 | 603.56 | 1.5109 | 0.209383 |
| ZonaPort | 2 | 7.069 | 299 | 596.49 | 3.5347 | 0.029166 * |
| TipusDeFons | 4 | 223.235 | 295 | 373.26 | 55.8088 | < 2.2e-16 *** |

DPS (*Parapenaeus longirostris*)
 DPS - Abundance



Deviance Residuals:

| Min | 1Q | Median | 3Q | Max |
|----------|----------|----------|---------|---------|
| -2.37545 | -1.13508 | -0.49433 | 0.03751 | 2.28111 |

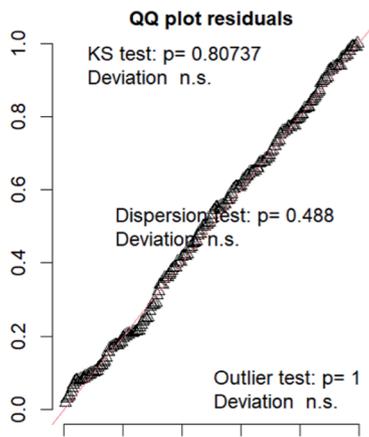
Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) |
|--------------------------------------|-----------|------------|---------|------------|
| (Intercept) | 7.261253 | 0.381225 | 19.047 | <2e-16 *** |
| Any2020 | 0.860502 | 0.313151 | 2.748 | 0.0060 ** |
| Any2021 | 1.012899 | 0.298831 | 3.390 | 0.0007 *** |
| EstacioSpring | -0.405986 | 0.353758 | -1.148 | 0.2511 |
| EstacioSummer | -0.591107 | 0.340811 | -1.734 | 0.0828 . |
| Estaciowinter | 0.249364 | 0.361672 | 0.689 | 0.4905 |
| ZonaPortNorth | -0.089212 | 0.302402 | -0.295 | 0.7680 |
| ZonaPortSouth | -0.296081 | 0.465391 | -0.636 | 0.5246 |
| TipusDeFonsMiddle continental shelf | 0.006816 | 0.549839 | 0.012 | 0.9901 |
| TipusDeFonsShallow continental shelf | -6.340271 | 0.567479 | -11.173 | <2e-16 *** |
| TipusDeFonsUpper slope | 0.282935 | 0.294407 | 0.961 | 0.3365 |

 Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

| | Df | Deviance | Resid. Dev | Df Resid. Dev | F | Pr(>F) |
|-------------|----|----------|------------|---------------|--------|-----------------------|
| NULL | | | | 223 | 352.18 | |
| Any | 2 | 7.158 | | 221 | 345.02 | 3.5789 0.02791 * |
| Estacio | 3 | 1.914 | | 218 | 343.10 | 0.6381 0.59038 |
| ZonaPort | 2 | 6.904 | | 216 | 336.20 | 3.4518 0.03169 * |
| TipusDeFons | 3 | 64.047 | | 213 | 272.15 | 21.3491 8.019e-14 *** |

DPS – Biomass



Deviance Residuals:

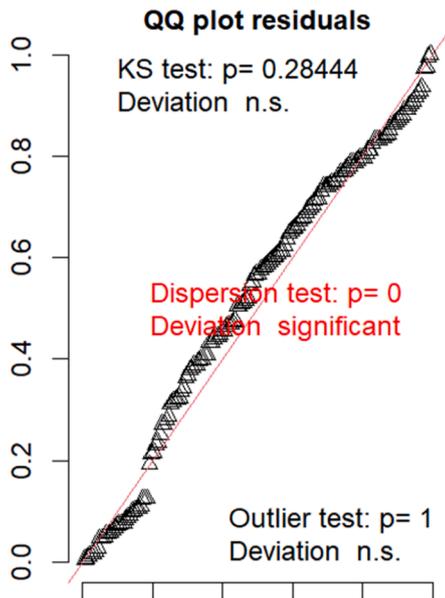
| Min | 1Q | Median | 3Q | Max |
|----------|----------|----------|---------|---------|
| -2.11697 | -1.21691 | -0.33352 | 0.07183 | 2.22252 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) | |
|--------------------------------------|----------|------------|---------|----------|-----|
| (Intercept) | 2.24790 | 0.30830 | 7.291 | 3.07e-13 | *** |
| Any2020 | 0.60738 | 0.25906 | 2.345 | 0.019049 | * |
| Any2021 | 0.86439 | 0.24687 | 3.501 | 0.000463 | *** |
| EstacioSpring | -0.19093 | 0.29032 | -0.658 | 0.510758 | |
| EstacioSummer | -0.21692 | 0.28231 | -0.768 | 0.442266 | |
| EstacioWinter | 0.05585 | 0.29864 | 0.187 | 0.851664 | |
| ZonaPortNorth | 0.10640 | 0.23874 | 0.446 | 0.655827 | |
| ZonaPortSouth | -0.07983 | 0.36958 | -0.216 | 0.828979 | |
| TipusDeFonsMiddle continental shelf | -0.02725 | 0.43738 | -0.062 | 0.950321 | |
| TipusDeFonsShallow continental shelf | -5.95250 | 1.22633 | -4.854 | 1.21e-06 | *** |
| TipusDeFonsUpper slope | 0.73888 | 0.23237 | 3.180 | 0.001474 | ** |

| | Df | Deviance | Resid. Df | Resid. Dev | F | Pr(>F) | |
|-------------|----|----------|-----------|------------|---------|-----------|-----|
| NULL | | | 223 | 350.74 | | | |
| Any | 2 | 11.058 | 221 | 339.69 | 5.5291 | 0.0039697 | ** |
| Estacio | 3 | 1.239 | 218 | 338.45 | 0.4131 | 0.7435792 | |
| ZonaPort | 2 | 18.010 | 216 | 320.44 | 9.0052 | 0.0001228 | *** |
| TipusDeFons | 3 | 67.680 | 213 | 252.76 | 22.5600 | 1.339e-14 | *** |

NEP (*Nephrops norvegicus*)
NEP – Abundance



Deviance Residuals:

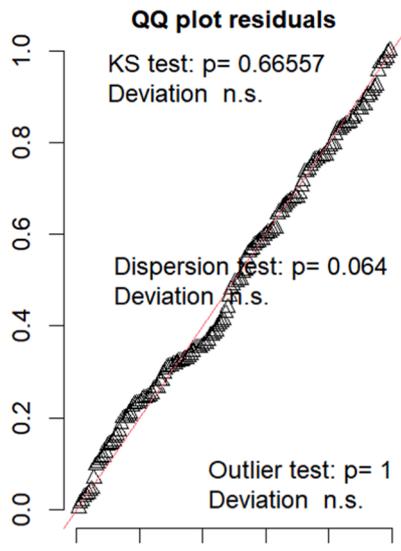
| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -2.1348 | -1.1307 | -0.4064 | 0.2111 | 3.3648 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) | |
|------------------------|----------|------------|---------|----------|-----|
| (Intercept) | 4.41853 | 0.36189 | 12.209 | <2e-16 | *** |
| Any2020 | -0.04122 | 0.32197 | -0.128 | 0.8981 | |
| Any2021 | -0.18183 | 0.31519 | -0.577 | 0.5640 | |
| EstacioSpring | -0.01763 | 0.36011 | -0.049 | 0.9610 | |
| EstacioSummer | 0.23411 | 0.35530 | 0.659 | 0.5100 | |
| EstacioWinter | 0.07757 | 0.37566 | 0.206 | 0.8364 | |
| ZonaPortNorth | 0.48969 | 0.26846 | 1.824 | 0.0681 | . |
| ZonaPortSouth | -0.13535 | 0.63739 | -0.212 | 0.8318 | |
| TipusDeFonsUpper slope | 2.99362 | 0.26185 | 11.433 | <2e-16 | *** |

| | Df | Deviance | Resid. Dev | Df | Resid. Dev | F | Pr(>F) |
|-------------|----|----------|------------|----|------------|---------|------------|
| NULL | | | 161 | | 304.42 | | |
| Any | 2 | 1.860 | 159 | | 302.56 | 0.9299 | 0.3946 |
| Estacio | 3 | 5.838 | 156 | | 296.72 | 1.9460 | 0.1198 |
| ZonaPort | 2 | 4.278 | 154 | | 292.44 | 2.1392 | 0.1178 |
| TipusDeFons | 1 | 94.530 | 153 | | 197.91 | 94.5304 | <2e-16 *** |

NEP – Biomass



Deviance Residuals:

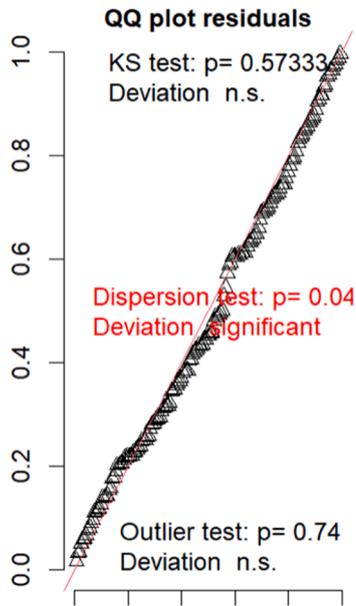
| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -2.4361 | -1.3921 | -0.4951 | 0.2631 | 4.4912 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) | |
|------------------------|----------|------------|---------|----------|-----|
| (Intercept) | 0.79684 | 0.27648 | 2.882 | 0.00395 | ** |
| Any2020 | -0.06802 | 0.24035 | -0.283 | 0.77716 | |
| Any2021 | -0.16964 | 0.23514 | -0.721 | 0.47063 | |
| EstacioSpring | 0.04397 | 0.26914 | 0.163 | 0.87023 | |
| EstacioSummer | 0.25512 | 0.26492 | 0.963 | 0.33554 | |
| EstacioWinter | 0.05120 | 0.28185 | 0.182 | 0.85584 | |
| ZonaPortNorth | 0.39170 | 0.20061 | 1.953 | 0.05087 | . |
| ZonaPortSouth | 0.10471 | 0.49544 | 0.211 | 0.83262 | |
| TipusDeFonsUpper slope | 2.84207 | 0.19637 | 14.473 | < 2e-16 | *** |

| | Df | Deviance | Resid. Dev | Df | Resid. Dev | F | Pr(>F) |
|-------------|----|----------|------------|----|------------|----------|---------------|
| NULL | | | 161 | | 383.87 | | |
| Any | 2 | 4.090 | 159 | | 379.78 | 2.0452 | 0.129360 |
| Estacio | 3 | 12.038 | 156 | | 367.74 | 4.0127 | 0.007254 ** |
| ZonaPort | 2 | 3.809 | 154 | | 363.93 | 1.9045 | 0.148890 |
| TipusDeFons | 1 | 168.622 | 153 | | 195.31 | 168.6219 | < 2.2e-16 *** |

ARA (*Aristeus antennatus*)
ARA – Abundance



Deviance Residuals:

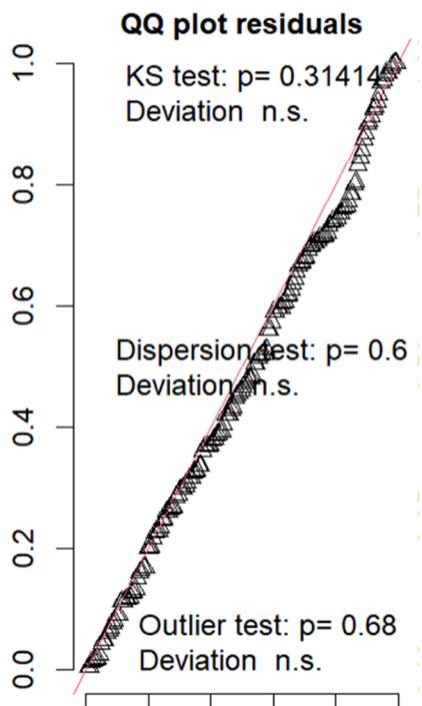
| Min | 1Q | Median | 3Q | Max |
|----------|----------|----------|----------|---------|
| -1.65595 | -1.20445 | -0.79044 | -0.07024 | 2.19567 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) |
|------------------------|----------|------------|---------|--------------|
| (Intercept) | 7.5889 | 0.4357 | 17.417 | < 2e-16 *** |
| Any2020 | 0.6946 | 0.3873 | 1.793 | 0.072946 . |
| Any2021 | -1.2606 | 0.3825 | -3.295 | 0.000983 *** |
| EstacioSpring | -0.4053 | 0.4360 | -0.930 | 0.352578 |
| EstacioSummer | 0.3028 | 0.4302 | 0.704 | 0.481560 |
| EstacioWinter | -0.2256 | 0.4540 | -0.497 | 0.619175 |
| ZonaPortNorth | 2.0436 | 0.3260 | 6.269 | 3.64e-10 *** |
| ZonaPortSouth | 2.5018 | 0.7541 | 3.318 | 0.000908 *** |
| TipusDeFonsUpper slope | -6.8163 | 0.3171 | -21.494 | < 2e-16 *** |

| | Df | Deviance | Resid. Dev | Df | Resid. Dev | F | Pr(>F) |
|-------------|----|----------|------------|-----|------------|--------|--------|
| NULL | | | 377.26 | 161 | | | |
| Any | 2 | 0.707 | 376.55 | 159 | | 0.3536 | 0.7022 |
| Estacio | 3 | 1.031 | 375.52 | 156 | | 0.3438 | 0.7936 |
| ZonaPort | 2 | 2.362 | 373.16 | 154 | | 1.1809 | 0.3070 |
| TipusDeFons | 1 | 208.852 | 164.30 | 153 | 208.8516 | <2e-16 | *** |

ARA – Biomass



Deviance Residuals:

| Min | 1Q | Median | 3Q | Max |
|---------|---------|---------|--------|--------|
| -2.8733 | -0.6669 | -0.5399 | 0.1203 | 4.3948 |

Coefficients:

| | Estimate | Std. Error | z value | Pr(> z) | |
|------------------------|----------|------------|---------|----------|-----|
| (Intercept) | 3.87104 | 0.15197 | 25.473 | < 2e-16 | *** |
| Any2020 | -0.01914 | 0.14148 | -0.135 | 0.89239 | |
| Any2021 | -0.27715 | 0.14182 | -1.954 | 0.05067 | . |
| EstacioSpring | 0.42935 | 0.16238 | 2.644 | 0.00819 | ** |
| EstacioSummer | 0.46131 | 0.15716 | 2.935 | 0.00333 | ** |
| EstacioWinter | 0.36378 | 0.16679 | 2.181 | 0.02918 | * |
| ZonaPortNorth | 0.27225 | 0.12162 | 2.239 | 0.02519 | * |
| ZonaPortSouth | 0.63203 | 0.22632 | 2.793 | 0.00523 | ** |
| TipusDeFonsUpper slope | -5.96102 | 0.27504 | -21.674 | < 2e-16 | *** |

| | Df | Deviance | Resid. Df | Resid. Dev | F | Pr(>F) | |
|-------------|----|----------|-----------|------------|-----------|-----------|-----|
| NULL | | | 161 | 1635.72 | | | |
| Any | 2 | 10.26 | 159 | 1625.46 | 5.1295 | 0.005919 | ** |
| Estacio | 3 | 23.25 | 156 | 1602.21 | 7.7492 | 3.586e-05 | *** |
| ZonaPort | 2 | 36.41 | 154 | 1565.80 | 18.2070 | 1.238e-08 | *** |
| TipusDeFons | 1 | 1417.62 | 153 | 148.18 | 1417.6181 | < 2.2e-16 | *** |

Appendix A. Supporting information

Supplementary data associated with this article can be found in the online version at [doi:10.1016/j.marpol.2024.106512](https://doi.org/10.1016/j.marpol.2024.106512).

Data availability

Data will be made available on request.

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